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Coral Uptake of Inorganic Phosphorus and Nitrogen Negatively Affected by Simultaneous Changes in Temperature and pH

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Abstract

The effects of ocean acidification and elevated seawater temperature on coral calcification and photosynthesis have been extensively investigated over the last two decades, whereas they are still unknown on nutrient uptake, despite their importance for coral energetics. We therefore studied the separate and combined impacts of increases in temperature and $p\text{CO}_2$ on phosphate, ammonium, and nitrate uptake rates by the scleractinian coral *S. pistillata*. Three experiments were performed, during 10 days i) at three pH_T conditions (8.1, 7.8, and 7.5) and normal temperature (26°C), ii) at three temperature conditions (26°, 29°C, and 33°C) and normal pH_T (8.1), and iii) at three pH_T conditions (8.1, 7.8, and 7.5) and elevated temperature (33°C). After 10 days of incubation, corals had not bleached, as protein, chlorophyll, and zooxanthellae contents were the same in all treatments. However, photosynthetic rates significantly decreased at 33°C, and were further reduced for the pH_T 7.5. The photosynthetic efficiency of PSII was only decreased by elevated temperature. Nutrient uptake rates were not affected by a change in pH alone. Conversely, elevated temperature (33°C) alone induced an increase in phosphate uptake but a severe decrease in nitrate and ammonium uptake rates, even leading to a release of nitrogen into seawater. Combination of high temperature (33°C) and low pH_T (7.5) resulted in a significant decrease in phosphate and nitrate uptake rates compared to control corals (26°C, pH_T = 8.1). These results indicate that both inorganic nitrogen and phosphorus metabolism may be negatively affected by the cumulative effects of ocean warming and acidification.

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Introduction

Ocean acidification is the result of anthropogenic carbon dioxide (CO_2) emissions partially dissolving into seawater and progressively declining its pH [1]: over the 20th century, the oceans' average pH_T (total scale) has decreased by 0.1 unit from 8.21 to 8.10 [2,3], and it is predicted to further decrease by 0.3–0.5 unit by the end of this century [2,4]. The majority of calcifying organisms, and particularly scleractinian corals, are negatively affected by ocean acidification, as shown by the decrease in calcification, which is one of the main processes, with photosynthesis, studied up to now in this context. A drop in pH is indeed known to affect the carbonate cycle [4–6], reducing carbonate ions that corals use to build their skeleton, and leading to reduced coral calcification rates [1,7–10].

Beyond its impact on the carbonate cycle, ocean acidification also alters other elemental cycles, such as those of nitrogen and phosphorus [11–14]. Both nutrients are however essential for coral metabolism. Indeed, reef-building corals are living in nutrient-poor tropical waters, where the supply of available nutrient sources (zooplankton, dissolved and particulate organic matter, inorganic nutrients) is generally low [15]. Yet, by an efficient nutrient

recycling between reef biota, coral reefs retain a high productivity. Corals have for instance adapted to their oligotrophic environment by developing a symbiosis with dinoflagellates of the genus *Symbiodinium*, commonly called zooxanthellae. These symbionts largely contribute to the nutrition of their animal host by providing 1) photosynthesis-derived carbon to the animal tissue [16]; and 2) essential nutrients, such as nitrogen and phosphorus, either directly taken up from the external environment or recycled from the animal wastes [17]. These nutrients are combined to the products of photosynthesis and are transferred back to the host, mainly in the form of essential amino acids for nitrogen [18]. It has been calculated that uptake of inorganic nitrogen, at natural concentrations, contributes approximately 30% to the daily nitrogen requirement of the species *Acropora palmata* for gamete and mucus production, growth, and tissue repair [19]. In another coral species, *Pocillopora damicornis*, uptake of ammonium could even completely satisfy the nitrogen demand of this coral at field concentrations [20]. Concerning phosphorus, it enters into the composition of many biological molecules (DNA, RNA, phospholipids) and has a role in several biochemical mechanisms (through ATP). It controls in part coral growth and zooxanthellae photosynthesis [21,22]. The issue of nutrient limitation of corals

and their symbionts is therefore of prime interest, as nutrient provision sustains corals' metabolism in general, and in particular during thermal bleaching events. It has indeed been shown that nutrient-repleted corals, with higher zooxanthellae density and photosynthetic rates, are less susceptible to bleaching or nutrient shortage [23,24]. Provision of nutrient to corals is also important for the entire reef, because healthy corals sustain a high reef biomass production.

Despite the importance of nutrient provision for coral physiology and coral reef functioning, and its potential alteration by ocean acidification, only three studies have reported on the impacts of pH on nutrient uptake by symbionts or symbiotic cnidarians, and none of them have considered these effects in the view of ocean acidification. Rahav et al. [25] showed that the activity of glutamate dehydrogenase, an enzyme involved in the assimilation of ammonium, was higher at pH_T 7.3 than at pH_T 8.1 in the coral *S. pistillata*. D'Elia et al. [26] reported that ammonium uptake was insensitive to a pH_T decrease between 8.8 and 7.8 in zooxanthellae freshly isolated from the giant clam *Tridacna crocea*. A final study showed that ammonium accumulated more in the hemolymph of the giant clam *Tridacna gigas* at pH_T 7.4 than at pH_T 7.9 [27]. Therefore, the first aim of our study was to test if a decrease in external pH changed the uptake rates of inorganic nitrogen and phosphorus by the scleractinian coral *S. pistillata*. The impacts might be due to changes in the coral metabolism with acidification, which would in turn i) modify nutrient requirements as known in phytoplankton [13,28] and higher plants [29–31], or ii) decrease the energy expended for nutrient uptake, as may be the case if more energy is used for calcification under acidified conditions [32,33]. The impacts may also be related to changes in the relative abundances of the protonated species of nitrogen and phosphorus. Indeed, phosphate is transported via a sodium/phosphate symporter in the coral *Stylophora pistillata* [34], but the ability of this transporter to discriminate among the protonated species is still unknown. For nitrogen, no studies to date have elucidated the transporters involved in ammonium and nitrate uptake in corals. However, the potential impact of speciation changes on nutrient transport cannot be neglected.

Besides oceanic acidification, anthropogenic CO_2 emissions have also driven an increase in the oceans' average temperature by $0.4\text{--}1.0^\circ\text{C}$ in the past four decades [2,35]. Seawater temperatures are predicted to further rise, and cause several mass coral mortalities over the century [36,37]. Indeed, as tropical corals usually live close to their upper thermal limit, they are very sensitive to exposure to elevated temperature [36,38–40]. One of the first responses of corals to thermal stress is bleaching, the disruption of the symbiotic association and the expulsion of the zooxanthellae [36,40–43], which in turn affects coral calcification [7,9,44,45] and photosynthesis [9,41,46–48]. As for acidification, the impacts of elevated seawater temperature on the above two processes have been extensively studied in the last thirty years, but the effects on the nutrient requirements have never been examined. As temperature controls enzymatic activities, ocean warming could potentially impact nutrient uptake in corals. The second aim of this study was therefore to test the effects of a thermal stress on the uptake rates of inorganic nitrogen and phosphorus by the coral *S. pistillata*. Finally, we also tested both effects of oceanic acidification and elevated temperature on these uptake rates, since the two stressors are likely to act simultaneously on reefs in the future [48].

The general aim of the present study was to test, for the first time, the effects of increases in temperature and pCO_2 of magnitudes similar to those expected by 2100 and beyond on the nutrient uptake rates of the scleractinian coral *S. pistillata*. This

study thus enriches our existing predictive capabilities on corals' response to future changes.

Methods

Organisms and culture conditions

Colonies of the zooxanthellate coral *S. pistillata* were initially obtained from the Red Sea. Methods of collection were reviewed and approved by CITES under the permit number DCI/89/32, and importation was performed under CITES permit number 125/SPV. Colonies were then cultured at the Centre Scientifique de Monaco (CSM), under controlled conditions (26°C , salinity of 38). 1.5 month before the start of the experiment, the ca. 220 nubbins (i.e. branch tips) used during the experiment were prepared by cutting branches (2.5 ± 1.0 cm long and 0.6 ± 0.3 cm in diameter) of about twenty parent colonies with pliers after Tambutté et al. [49]. Nubbins were attached on nylon wires and suspended in aquaria until tissues fully covered the skeleton. They were lightly fed once a week with *Artemia salina* nauplii. In order to avoid the additional and undesired effects of bleaching on the impacts of a pH- and/or a temperature-stress, corals were maintained under a low photosynthetic active radiation (PAR) of $110 \mu\text{mol m}^{-2} \text{s}^{-1}$, using metal halid lamps (Philips, HPIT, 400W, Eindhoven, The Netherlands), with a photoperiod of 12h:12h light:dark, and measured using a spherical quantum sensor (LiCor LI-193), with a 12 h:12 h dark:light cycle. Nubbins were then randomly assigned to their experimental 40-L tanks and allowed to acclimate for a week. Temperature was individually regulated in each aquarium at $26.0\pm 0.1^\circ\text{C}$ during the acclimation period, and pH was 8.1. Aquaria were continuously provided with unfiltered seawater, with a flow rate of ca. 40 L day^{-1} .

Experimental design

To examine the effects of exposure to lowered pH and/or elevated temperatures on the uptake rates of ammonium, nitrate, and phosphate by *S. pistillata*, three experiments were performed, in which nubbins were incubated for 10 days i) at three pH_T conditions (8.1, 7.8, and 7.5) at normal temperature (26°C), ii) at three temperature conditions (26° , 29°C , and 33°C) at normal pH_T (8.1), and iii) at three pH_T conditions (8.1, 7.8, and 7.5) at elevated temperature (33°C). A short-term exposure of 10 days was chosen to be able to compare our results with previous ones. Indeed, it is usually the length of a thermal stress above 30°C , either observed *in-situ* [50] or applied in laboratory experiments [23,51,52]. A longer thermal stress generally induces a complete bleaching of the corals (and therefore a complete change in their physiology) and might lead to their death. A pCO_2 stress of 10 days also ranges in the mean culture length of previous experiments (mean of 12 days in the comprehensive Table 2 from Erez et al. 2011 if studies lasting over a c.a. year are excluded) [33].

Corals were not fed during the 10-day incubations, in order to minimize the impact of organic nutrients on uptake rates. To avoid any undesirable "tank" effects, the aquaria were carefully cleaned once a week to minimize algal growth on the walls and nylon wires. Salinity and irradiance were also monitored during the course of the experiments. This maintenance ensured that similar conditions prevailed in all the tanks, except for the fixed parameters (pH and seawater temperature). The three experiments were set up using 3 conditions with duplicated aquaria.

In experiment i), aquaria were maintained at $26\pm 0.2^\circ\text{C}$ and 3 different pHs (a normal pH_T : 8.09 ± 0.04 , i.e. $378 \mu\text{atm CO}_2$; a pH_T level projected for the end of the century: 7.78 ± 0.06 , i.e. $903 \mu\text{atm CO}_2$; and a very low pH_T level: 7.46 ± 0.04 , i.e. $2039 \mu\text{atm CO}_2$). The pH was controlled using a pH-stat system

(IKS, Karlsbad, accuracy ± 0.05 pH_T unit) by bubbling independently pure CO_2 in each tank that was continuously aerated with CO_2 -free air. A temperature of 26°C was kept constant inside each aquarium using heaters connected to electronic controllers (Biotherm, $\pm 0.2^\circ\text{C}$ accuracy) and was logged at 10-min intervals using individual temperature recorders (Seamon). Corals were maintained 10 days under these conditions before their nutrient uptake rates were measured.

In experiment ii), aquaria were maintained at the control pH_T (8.1) and at 3 different temperatures (a normal temperature: $26.0 \pm 0.2^\circ\text{C}$, the temperature projected for the end of the century: $29.0 \pm 0.2^\circ\text{C}$; and a very high temperature: $33.0 \pm 0.2^\circ\text{C}$). Temperature, salinity and irradiance were also repeatedly monitored over the course of the experiment, and the maintenance procedure was the same as in experiment i). Corals were also maintained 10 days under these conditions before the nutrient uptake measurements. Before the beginning of the 10-days incubation period, for the two tanks in which temperature was increased to 33°C , corals were first acclimated for 7 days at 29°C then 2 days at 31°C and 2 additional days at 33°C . This gradual temperature increase over 11 days prevented any thermal shock and coral mortality.

Finally, experiment iii) was performed using the same design as in experiment i), except that the temperature was set to 33°C (using the same gradual procedure as in experiment ii) for all the aquaria).

Data of seawater temperature and carbon chemistry were measured as in Houlbrèque et al. [53] and are presented in Table 1. Briefly, three 20 mL samples of seawater were collected daily in each tank, filtered through $0.45 \mu\text{m}$ GF/F Whatman filters, poisoned with 0.05 ml of 50% HgCl_2 to avoid biological alteration, and finally stored in glass bottles in the dark at 4°C . pH_T was measured using a Metrohm, 826pH meter, equipped with a glass electrode calibrated on the total scale using Tris/HCl and 2-aminopyridine/HCl buffer solutions with a salinity of 38 [54]. Means pH_T were calculated from hydrogen ion concentrations of each measurement and then re-converted back to pH [54]. These measurements were used to adjust every day the pH values of the pH-stat system. Total alkalinity (TA) was calculated from the Gran function. Titrations of TA standards were within $0.7 \mu\text{mol kg}^{-1}$ of the nominal value. Mean TA of seawater was $2508 \pm 16 \mu\text{mol kg}^{-1}$ and remained stable during the whole experiment. pCO_2 was calculated from pH_T , TA, temperature and salinity using the free-access CO_2 Systat package. For each experimental treatment, the parameters of the carbonate system remained constant over the 10-days exposure periods (repeated measures ANOVA, all $p > 0.05$).

Measurements of nutrient uptake rates

After 10 days of incubation, inorganic nutrient uptake was studied by following the depletion of phosphate, ammonium, or nitrate over time in beakers containing filtered-seawater enriched with the corresponding nutrient. For each nutrient (PO_4 , NH_4 , or NO_3), six 200-mL beakers were maintained in a water bath at the pH, temperature and irradiance to which the nubbins were previously exposed. Five randomly-selected coral nubbins per experimental condition, each suspended to a nylon wire, were introduced in separate beakers. The last beaker, without coral, served as a control for changes in nutrient concentrations due to adsorption onto beaker surface, to consumption by microbial activity, or to air contamination (particularly for ammonium). After a 10–30 min acclimatization period, beakers were enriched with $300 \mu\text{L}$ of a 10-mmol L^{-1} buffer solution of KH_2PO_4 , NH_4Cl , or KNO_3 to reach an initial concentration of $3 \text{ } \mu\text{mol L}^{-1}$. This concentration was chosen because it corresponded to the

plateau in previous concentration-dependent uptake experiments with the same coral species [34,55]. Magnetic stirring bars ensured proper homogenization of nutrient inside the beakers. A 10-mL water sample was taken from each beaker immediately after enrichment and then every 15 min for 60 min. 10-mL samples were also taken in the beakers at the beginning and at the end of the incubation in order to verify that pH had not changed during that period. pH values were measured on total scale (pH_T) with a glass electrode connected to a pH meter (Metrohm, 826 pH mobile) calibrated on the total scale using Tris/HCl and 2-aminopyridine/HCl buffers with a salinity of 38 [54].

Ammonium and phosphate levels were determined manually, according to the spectrofluorimetric method of Holmes et al. [56], and using the ascorbic acid technique of Murphy and Riley [57], respectively. Levels of nitrate were determined using an autoanalyzer (Axflow) according to Aminot and K  rouel [58]. Uptake rates of ammonium, nitrate, and phosphate were calculated as the quantity taken up by each nubbin in 60 min, after correction of concentrations for the diminution of the beakers' volume and after having verified that the uptake rates were linear during the first hour. Uptake rates were normalized to the total chlorophyll content, to the zooxanthellae concentration, or to the surface area of the nubbins (see below).

Measurements of physiological parameters

Photosynthesis and respiration were measured at the end of the experiments, for 3 randomly selected corals per experimental condition, using the same method as described in Godinot et al. [55]. Seawater in the respirometric chambers was at the same pH and temperature as the corresponding experimental tanks. Photosynthesis was measured at a PAR of 110 and $300 \mu\text{mol m}^{-2} \text{ s}^{-1}$. Data were normalized to the surface area of the nubbins and to their chlorophyll content (see below). Additionally, we measured the maximum quantum yield of photosystem II (F_v/F_m of PSII) and the electron transport rate of PSII ($\text{ETR} = \text{dark adapted } F_v/F_m \times 0.5 \times \text{PAR}$) at 9 different light intensities ($0 - 3000 \mu\text{mol m}^{-2} \text{ s}^{-1}$ PAR) at the end of the two experiments, for 5 randomly selected corals per experimental condition, using pulse-amplitude-modulated chlorophyll fluorometry (diving PAM, Waltz, Germany).

Chlorophyll (chl), protein (prot), and zooxanthellae (zoox) concentrations were measured at the end of the experiments, on all nubbins used for the uptake rate measurements (15 nubbins per experimental condition). Tissues and zooxanthellae were striped from the skeleton in 10 mL of filtered seawater using an Air-Pik, then homogenized using a Potter grinder, and divided into 3 aliquots for chlorophyll, zooxanthellae, and protein measurements. Chlorophyll *a* and *c2* were extracted in 99% acetone (24 h at 4°C). The extracts were centrifuged at $11,000 \times g$ for 10 min at 4°C , and the absorbances measured at 630, 663, and 750 nm. Concentrations were computed according to the 100% acetone equations of Jeffrey and Humphrey [59]. Zooxanthellae were counted using a counting chamber and an improved version of the Histolab v.5.2.3 image analysis software (Microvision, Every, France). Proteins were extracted in 0.5 mol L^{-1} NaOH (60°C , 5 h) and measured using the commercially available BC Assay Interchim kit, and results analyzed with the GENESIS software (v.3.3). Results were normalized to the surface area of nubbins, measured with the wax technique [60].

Statistical analyses

Equality of variances and normality of residuals were tested using Levene and Shapiro-Wilk tests (Statgraphics Centurion version 15). Data from the tanks replicated between the 3

Table 1. Carbonate system parameters during the experiments.

Experiment	Temperature (°C)	pH _T	pCO ₂ (μatm)	HCO ₃ ⁻ (μmol kg ⁻¹)	CO ₃ ²⁻ (μmol kg ⁻¹)	CO ₂ (μmol kg ⁻¹)	DIC (μmol kg ⁻¹)
1) Effect of pH	26.0±0.2	8.09±0.04	378±15	1890±4	267.7±2.0	55.2±2.0	2213±2.2
	26.0±0.2	7.78±0.06	903±51	2220±48	150.1±4.7	25.1±1.2	2395±16.7
	26.0±0.2	7.46±0.04	2039±82	2345±6	80.6±2.6	11.8±0.2	2437±5.5
2) Effect of temperature	26.0±0.2	8.08±0.02	375±21	1885±10	268.7±1.6	54.7±2.2	2208±6.0
	29.0±0.2	8.09±0.01	405±17	1891±4	270.2±2.7	59.1±1.8	2220±3.1
	33.0±0.2	8.07±0.04	429±34	1878±8	269.5±1.9	62.6±1.5	2210±4.2
3) Effect of pH x temperature	33.0±0.2	8.08±0.01	434±25	1892±5	269.3±2.1	63.3±1.7	2225±3.9
	33.0±0.2	7.76±0.02	1037±42	2210±34	155.4±5.3	28.9±1.5	2394±14.8
	33.0±0.2	7.45±0.01	2242±75	2337±10	82.8±2.9	12.9±0.1	2433±4.5

Temperature and parameters of the carbonate system in each treatment during a 10-day exposure to lowered pH and/or elevated temperatures. The values reported are averages ± SE of n=60 measurements performed during the 10-days exposure periods. The total alkalinity was constant and equal to 2508±16 imol kg⁻¹. pCO₂ was calculated from pH_T, TA, temperature, and salinity using the free-access CO₂ Systat package.

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experiments (26°C or 33°C, and pH_T of 8.1) were compared using unpaired *t*-tests. Since there was no significant difference between replicated tanks, results from those tanks were pooled together for the subsequent statistical analyses. Effects of temperature and pH on nutrient uptake rates, areal chlorophyll, zooxanthellae and protein content, photosynthesis, respiration, F_v/F_m, and ETR_{max} were examined using analyses of variance tests (2-ways ANOVAs, with temperature and pH as fixed factors; StatView version 5.0). When significant differences were found, ANOVAs were followed by protected least significant difference (PLSD) Fisher post hoc tests to attribute differences between specific treatments. The effect of temperature alone on the above-cited parameters was tested using a 1-way ANOVA.

Results

Effect of lowered pH or/and elevated temperature on physiological parameters

After 10 days of incubation, there was no effect of either pH, temperature, or both on the areal content of zooxanthellae (Fig. 1a,d,g), chlorophyll (Fig. 1b,e,h), and protein (Fig. 1c,f,i), nor on the rates of respiration normalized to the chlorophyll content (Fig. 2; 1-way and 2-ways ANOVA, all *p*>0.05). The rate of photosynthesis normalized to the chlorophyll content was not affected by a change in pH at the control temperature of 26°C (Fig. 2a), but was impacted by a change in temperature alone (Fig. 2b) or combined with a low pH (Fig. 2c). Therefore, at pH_T=8.1, photosynthesis was ca. 2 times lower at 33°C than at 26°C or 29°C (2-ways ANOVA, *p*=0.0004, *F*₂=17.51). At 33°C, photosynthesis was also 1.8 times lower for pH_T=7.5 than for pH_T=7.8 and 8.1 (2-ways ANOVA, *p*=0.038, *F*₂=4.48). There was however no significant interaction of pH and temperature on these rates of photosynthesis (2-ways ANOVA, *p*=0.28, *F*₂=1.37 for an irradiance of 300 μmol m⁻² s⁻¹). Results were the same when respiration and photosynthetic rates were normalized to the skeletal surface area (cm²). There was no effect of pH on F_v/F_m and ETR, both at 26°C and 33°C (Table 2 for F_v/F_m, and Fig. 3a,c for ETR; 2-way ANOVAs, *p*=0.93, *F*₂=0.07 for F_v/F_m; and *p*=0.87, *F*₂=0.14 for ETR). Conversely, temperature had an effect at pH_T=8.1 (Table 2 for F_v/F_m, and Fig. 3b for ETR; 1-

way ANOVAs, *p*=0.01, *F*₂=7.52 for F_v/F_m; *p*<0.0001, *F*₂=24.70 for ETR): F_v/F_m and ETR were significantly higher at 29°C than at 26°C (PLSD Fisher tests, both *p*<0.02, *df*=9), and they were significantly lower at 33°C than at 26°C (PLSD Fisher tests, both *p*<0.001, *df*=9). There was however no significant interaction of pH and temperature on these parameters (2-ways ANOVAs, *p*=0.67, *F*₂=0.40 for F_v/F_m; *p*=0.51, *F*₂=0.69 for ETR).

Effect of lowered pH or/and elevated temperature on nutrient uptake rates

Results of the statistical tests were the same regardless of the normalization used for uptake rates. Uptake data in this paper are presented as normalized to the chl content.

A decrease in pH at 26°C (Fig. 4a,b,c) did not significantly affect ammonium (2-ways ANOVA, *p*=0.29, *F*₂=1.35), nitrate (2-ways ANOVA, *p*=0.39, *F*₂=1.03), and phosphate (2-ways ANOVA, *p*=0.11, *F*₂=2.49) uptake rates.

Conversely, elevated temperatures had a significant effect on nutrient uptake. Indeed, under a normal pH_T of 8.1 (Fig. 4d,e,f), ammonium uptake rates first increased by 5 fold at 29°C compared to 26°C (PLSD test, *p*=0.004), but then severely decreased to negative values (i.e. release of inorganic nitrogen) at 33°C (1-way ANOVA, *p*=0.01, *F*₂=6.39). An equivalent decrease to negative values at 33°C was also observed for nitrate (1-way ANOVA, *p*=0.05, *F*₂=3.63), while phosphate uptake rates almost doubled at 33°C (1-way ANOVA, *p*=0.003, *F*₂=9.62).

At 33°C and a low pH_T of 7.5 (Fig. 4g,h,i), ammonium uptake rates were the same as at pH_T of 8.1 and temperature of 26°C (Fig. 4a,d), and there was no significant interaction of pH and temperature (2-ways ANOVA, *p*=0.10, *F*₂=2.56). On the contrary, corals excreted nitrate at 33°C and a low pH_T of 7.5, and release rates did not vary with pH (2-ways ANOVA, *p*=0.88, *F*₂=0.14). Conversely, phosphate uptake rates severely decreased at 33°C and a low pH_T of 7.5, by 4 fold when compared to 33°C and a normal pH_T of 8.1 (2-ways ANOVA, *p*<0.0001, *F*₂=27.21), or by 2 fold when compared to control corals (pH_T=8.1, 26°C; 2-ways ANOVA, *p*<0.0001, *F*₂=22.78). There was a significant interaction of pH and temperature (2-ways ANOVA, *p*<0.0001, *F*₂=20.56) on phosphate uptake rates.

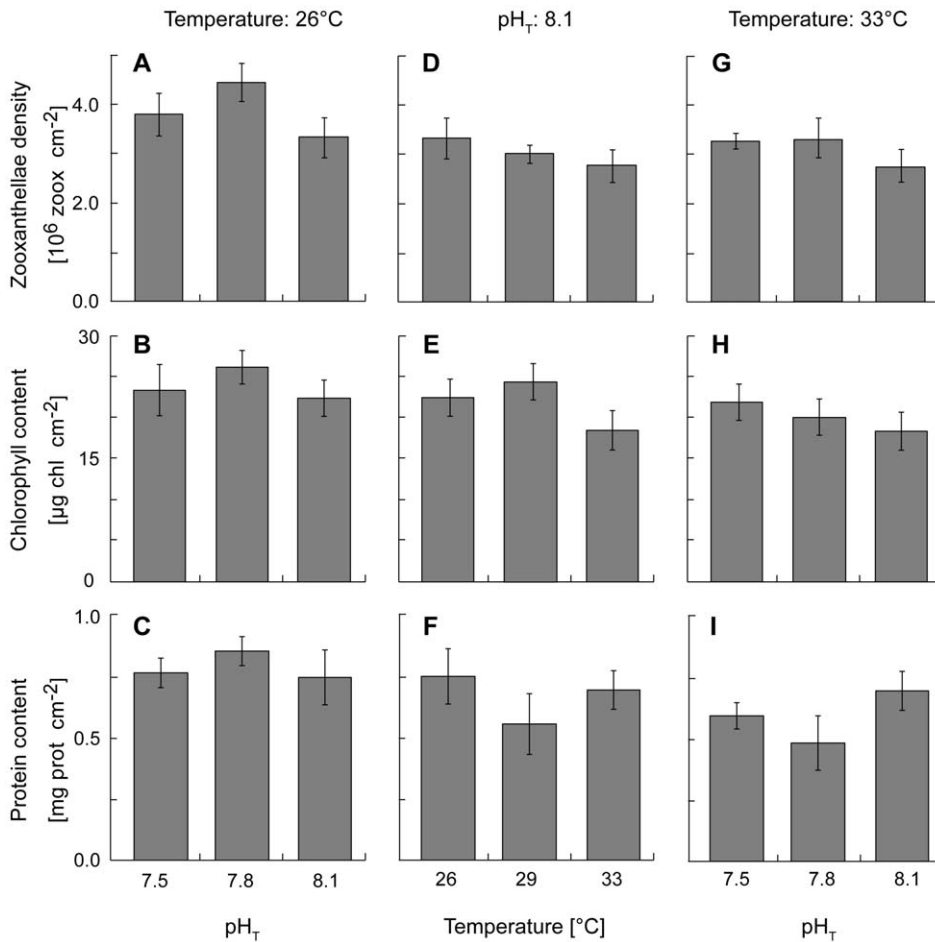


Figure 1. Effect of pH and/or temperature on coral biomass. Effect of a 10-day exposure to lowered pH and/or elevated temperatures on zooxanthellae (A, D, G), chlorophyll (B, E, H), and protein (C, F, I) content of *Stylophora pistillata* nubbins. Corals were incubated either under 3 different pH_T at 26°C (A, B, C), under 3 different temperatures at $pH_T=8.1$ (D, E, F), or under 3 different pH_T at 33°C (G, H, I). Data are presented as mean \pm SE of 15 nubbins per treatment. doi:10.1371/journal.pone.0025024.g001

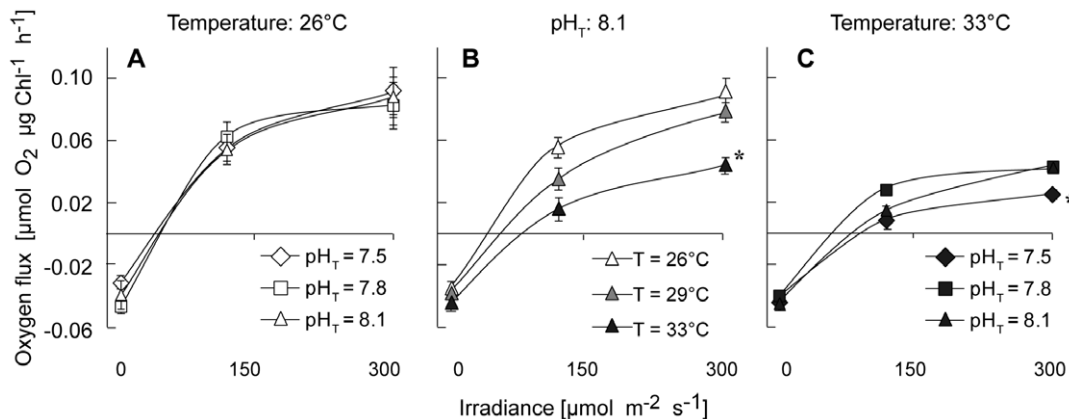


Figure 2. Effect of pH and/or temperature on coral photosynthesis and respiration. Effect of a 10-day exposure to lowered pH and/or elevated temperatures on oxygen fluxes in *Stylophora pistillata* nubbins. Corals were incubated either under 3 different pH_T at 26°C (A), under 3 different temperatures at $pH_T=8.1$ (B), or under 3 different pH_T at 33°C (C). Data are presented as mean \pm SE of 3 nubbins per treatment. Lozenges: $pH_T=7.5$, squares: $pH_T=7.8$, triangles: $pH_T=8.1$, open symbols: $T=26^\circ\text{C}$, gray symbols: $T=29^\circ\text{C}$, dark symbols: $T=33^\circ\text{C}$. Stars indicate treatments significantly different from the control ($pH_T=8.1$ for A and C; $T=26^\circ\text{C}$ for B). doi:10.1371/journal.pone.0025024.g002

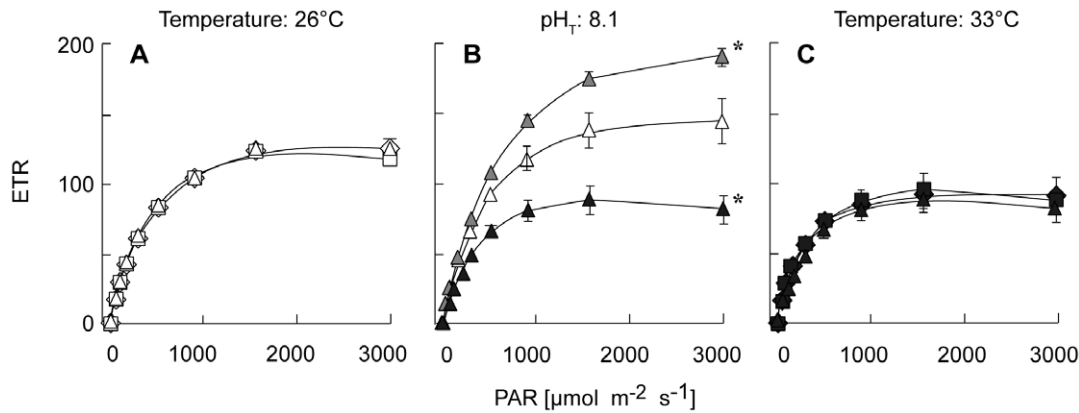


Figure 3. Effect of pH and/or temperature on coral PSII photosynthetic activity. Effect of a 10-day exposure to lowered pH and/or elevated temperatures on the electron transport rate (ETR) of *Stylophora pistillata* nubbins. Corals were incubated either under 3 different pH_T at 26°C (A), under 3 different temperatures at pH_T=8.1 (B), or under 3 different pH_T at 33°C (C). Data are presented as mean ± SE of 5 nubbins per treatment. Symbols are the same as on Fig. 2. Stars indicate treatments significantly different from the control (pH_T=8.1 for A and C; T=26°C for B). doi:10.1371/journal.pone.0025024.g003

Discussion

This study presents the changes in the uptake of inorganic phosphorus and nitrogen by a scleractinian coral, *S. pistillata*, under elevated temperature and/or decreased pH, thus bringing new insights into the ability of corals to respond to changes in their environment. To the best of our knowledge, the effects of acidification and/or elevated temperature on nutrient uptake by cnidarians, or even phytoplankton species, have been poorly investigated up to date. We showed that a short-term seawater acidification alone had no impact on the nutrient uptake rates, while elevated temperature (33°C) increased the uptake of phosphate but severely decreased, and even reversed the uptake of nitrogen (nitrogen release). An elevation in both temperature and pCO₂ significantly decreased the uptake of phosphate and nitrate compared to normal conditions (pH_T 8.1 and 26°C), suggesting that overall, climate change will negatively affect nutrient supply in corals.

Effect of lowered pH or/and elevated temperature on physiological parameters

Even though corals did not bleach during the 10-days incubation under high temperature, photosynthesis was impacted, with an increase in the F_v/F_m and ETR at 29°C, and a decrease of the photosynthetic rates, F_v/F_m, and ETR at 33°C, as previously

observed [61]. A general decrease in photosynthesis has been observed when temperatures extend above 31°C [41,46,47,62,63], whereas a positive effect was reported when temperatures remained below 31°C [9,41,46,48,62,64]. The decrease in the rates of photosynthesis is often due to damages to the photosystems II (PSII) early during temperature stress [41,46,62,65,66], as observed in this study. Conversely to the temperature effect, there was no effect of pH alone on the photosynthetic rates, which is also in agreement with most of the previous studies [8,9,53,67–69]. This feature is attributed to the fact that corals do not rely solely on dissolved CO₂ for photosynthesis, but also largely depend on HCO₃⁻ [67,70–73]. Our results are however in disagreement with the study of Anthony et al. [48], who reported a decrease in the net productivity and an increase in the respiration of the corals *Acropora intermedia* and *Porites lobata* with seawater acidification even at normal temperature. However, the latter study used a higher natural irradiance (1000 μmol photons m⁻² s⁻¹) and a long-term exposure (8 weeks), which probably brought corals closer to their bleaching threshold. Finally, Hii et al. [74] reported that seawater acidification decreased photosynthesis in the coral *Galaxea fascicularis* and had no impact on that of *Porites cylindrica*, thus showing that the response of corals may be species-dependent.

Effect of lowered pH or/and elevated temperature on nutrient uptake rates

Maximum nutrient uptake rates correspond, as for all enzyme-involving processes, to an optimum in temperature and pH conditions [75,76]. This optimum can be different according to the nutrient or the organism considered. In this study, the optimum uptake rate for ammonium was achieved for the normal pH_T of 8.1 and a temperature of 29°C, while the highest phosphate uptake rate was observed for the normal pH_T of 8.1 and a temperature of 33°C. An increase in nitrate uptake rates was also observed at 29°C, although it was not significant. For both inorganic nitrogen and phosphorus, maximal rates were therefore achieved under elevated temperatures and at the current pH_T of 8.1.

Effect on ammonium uptake rates. The lack of impact of seawater pH alone (in the range studied) on ammonium uptake rates is in agreement with results reported on zooxanthellae freshly isolated from the symbiotic clam *Tridacna crocea*, for which no difference was found across a pH_T range of 7.8 to 8.8 [26]. On the

Table 2. Effect of pH and/or temperature on coral PSII photosynthetic efficiency.

pH _T	Temperature		
	26°C	29°C	33°C
7.5	0.65±0.002		0.62±0.021
7.8	0.65±0.004		0.63±0.009
8.1	0.66±0.015	0.69±0.002	0.63±0.011

Effect of a 10-day exposure to lowered pH and/or elevated temperatures on the maximum quantum yield of photosystem II (F_v/F_m of PSII) of *Stylophora pistillata* nubbins. Corals were incubated either under 3 different pH_T at 26°C, under 3 different temperatures at pH_T=8.1, or under 3 different pH_T at 33°C. Data are presented as mean ± SE of 5 to 10 nubbins per condition. doi:10.1371/journal.pone.0025024.t002

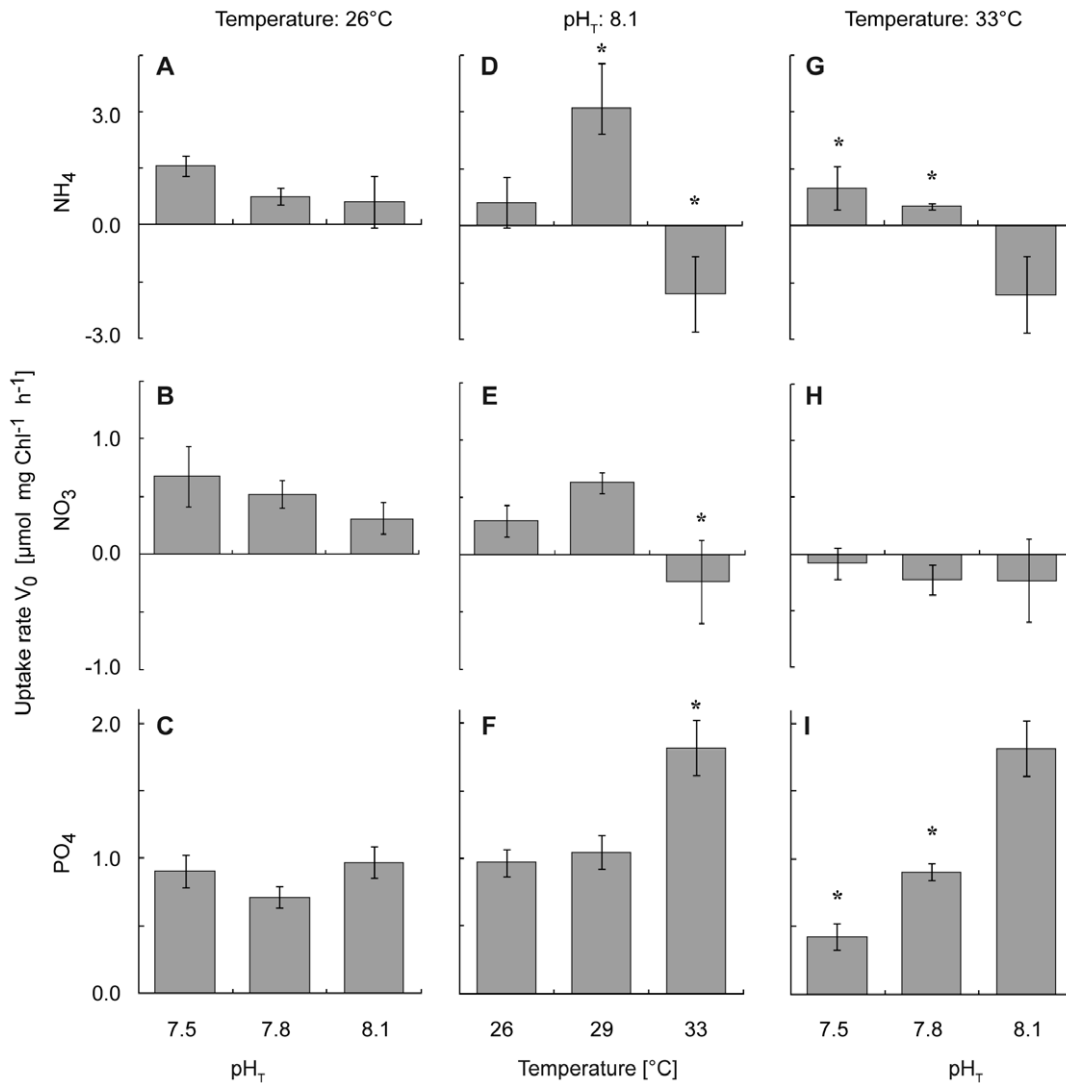


Figure 4. Effect of pH and/or temperature on coral nutrient uptake rates. Effect of a 10-day exposure to lowered pH and/or elevated temperatures on ammonium (A, D, G), nitrate (B, E, H), and phosphate (C, F, I) uptake rate of *S. pistillata* nubbins. Corals were incubated either under 3 different pH_T at 26°C (A, B, C), under 3 different temperatures at pH_T=8.1 (D, E, F), or under 3 different pH_T at 33°C (G, H, I). Data are presented as mean \pm SE of 5 nubbins per treatment. Stars indicate treatments significantly different from the control (pH_T=8.1 for A, B, C, G, H, I; T=26°C for D, E, F). doi:10.1371/journal.pone.0025024.g004

contrary, Fitt et al. [27] observed an inverse relationship between pH and ammonium accumulation in the hemolymph of the symbiotic clam *T. gigas*, over a pH_T range of 7.4 to 7.9. However, in the latter study, changes in pH did not occur in the surrounding seawater but directly inside the tissue, and were related to diel variations in zooxanthellae photosynthesis. In our study, ammonium uptake rates displayed a bell-shaped response to temperature-stress, with a 5-fold increase at 29°C and a net decrease, and even release, at 33°C. Since ammonium uptake is a light-stimulated process in symbiotic cnidarians [17,77–80] and is linked to zooxanthellae photosynthesis, the observed increase in uptake rates between 26°C and 29°C and the decrease between 29 and 33°C might be linked to the parallel increase and decrease in photosynthesis and photosynthetic efficiency. A similar release of ammonium was indeed observed in a previous study on the same coral species [25], when the photosynthetic chain was blocked by the electron transport inhibitor DCMU. In that latter study, excretion of ammonium was also elicited by the inhibition of the

glutamate synthase (GOGAT) [25], which could be an additional explanation for the observed excretion of NH₄⁺ at high temperature (33°C). Indeed, although the temperature-sensitivity of this enzyme is not known in corals, heat stress was shown to decrease its activity, as well as that of glutamine synthetase (GS), in higher plants [81]. The release of ammonium at 33°C was however no longer observed under low pH (pH_T=7.5), and the uptake rate measured under these conditions was comparable to the “control conditions” (pH_T=8.1, 26°C), therefore suggesting that CO₂ addition may completely offset the negative impact of temperature. One hypothesis to explain this result is that ammonium-assimilating enzymes (GS, GOGAT, and glutamate dehydrogenase GDH) [25,80,82–84] in corals are stimulated by seawater acidification, possibly through the impact of acidification on the intracellular pH [85]. For example, the activity of GS, GOGAT, and GDH are influenced by pH in an actinobacteria [86] and a cyanobacteria [87], with pH_T optima of 7.0 (GS), 7.2–7.6 (GOGAT), and 7.2–7.5 (GDH). In the coral *S. pistillata*, GDH

aminating activity is also pH-dependent, as it increased when pH_T was decreased between 8.1 and 7.3 [25]. However, such potential pH effects on the enzymatic activity were only significant at elevated temperature, although a positive but not significant trend was observed at 26°C. It is possible that the enzymatic activity was already optimal at 26°C, as suggested by the concordance of uptake rates measured in this study with those reported in the literature [17,26,34,79,88], and so acidification did not lead to a significant improvement of ammonium uptake at that temperature. On the contrary, when conditions were no longer optimal (i.e. thermal stress), acidification appears to have a significant alleviating effect. It is therefore possible that, as was demonstrated for calcification [9], the effects of high CO_2 do not manifest at low temperatures.

Effect on nitrate uptake rates. Concerning nitrate, the lack of impact of seawater acidification alone is similar to the response reported for freshwater phytoplankton [89], for pH_T s ranging between 6.6 and 9.2. Conversely, in some marine macroalgae, some authors [90–92] observed an enhancement of nitrate uptake at high CO_2 levels due to an increase in the activity of the nitrate reductase. Such enhancement may not have occurred in the coral *S. pistillata*, possibly because nitrate is not the main source of nitrogen [93,94] and therefore, the nitrate reductase might not be very active. As for ammonium, nitrate uptake rates were severely affected by a thermal stress at 33°C, therefore suggesting that the two nitrogen forms present the same temperature optimum for their uptake (i.e. 29°C), and that climate change in general may negatively impact the absorption of nitrogen by scleractinian corals. When seawater acidification was combined to elevated temperature, nitrate uptake rates did not decrease any further.

Effect on phosphate uptake rates. The response of phosphate uptake to changes in temperature and pH differed from the nitrogen response. While a thermal stress at 33°C decreased the uptake rates of nitrogen, it increased those of phosphate. The only previous study that examined the impact of seawater temperature on phosphate uptake by the tropical coral *Pocillopora capitata* used a low temperature range of 6 to 22°C [21], and showed that temperatures below 19°C decreased uptake rates. Enhancement of uptake rates by temperature suggests that, although phosphate uptake is light-stimulated [21,55,95,96], it may be less dependent on photosynthesis than ammonium uptake.

As for nitrogen, there was no pH effect alone on phosphate uptake rates. This lack of pH effect (in the same pH range), either on phosphate uptake rates or on the activity of the alkaline phosphatase, has previously been reported for marine phytoplankton [97] and macroalgae [90]. In corals, it is possible that, as for ammonium, the enzymatic activity involved in phosphate uptake and assimilation was optimal at 26°C regarding the metabolic condition of the corals, since the uptake rates were in agreement with rates reported in the literature [21,34,55]. When seawater acidification was combined to elevated temperature, uptake rates severely decreased, by 4 fold when compared to elevated temperature and normal pH, or by 2 fold when compared to control corals ($\text{pH}_T=8.1$, 26°C). These results therefore suggest that a combined thermal and acidification stress might impact the retrieval of phosphate by corals in the future. The temperature-dependent response of corals to acidification may be due to a higher/lower affinity of the carrier for the different phosphate species available under certain environmental conditions. At 33°C

and pH_T 7.5, the decrease in uptake rates may for instance have resulted from a higher affinity of the carrier towards the phosphate form less available due to more acidic conditions, i.e. the less abundant PO_4^{3-} . Indeed, as seawater pH_T decreases from 8.1 to 7.5, the relative abundance of HPO_4^{2-} increases from 82% to 92%, while PO_4^{3-} decreases from 18% to 5% (H_2PO_4^- remains negligible). This 3.6-fold decrease in the relative abundance of PO_4^{3-} corresponds to the observed 4-fold decrease in phosphate uptake rates. From previous work on *S. pistillata*, we showed that a sodium/phosphate symporter is involved in the uptake of phosphate in this coral [34]. Although the same type of symporter has been reported in a marine aplousia [98], no studies to date have examined the stoichiometry of such transporters in invertebrates. On the contrary, in vertebrates, three distinct families have been found, with a remarkable preference (i.e. higher affinity) for divalent HPO_4^{2-} in the type II family and for monovalent H_2PO_4^- in the type III family [99,100]. The possibility of preferences for specific phosphate forms therefore needs to be considered more extensively for the sodium/phosphate symporter of corals.

Conclusion

In conclusion, this study has shown a strong negative impact of increasing temperature on the ability of corals to take up inorganic nitrogen from the surrounding environment. High temperature indeed completely inhibited the uptake of either ammonium or nitrate, and even induced a nitrogen release from the corals. Although high temperature enhanced the uptake of inorganic phosphorus, corals cannot use this phosphorus without nitrogen, which is needed in all metabolic functions, such as tissue growth and repair, chlorophyll synthesis, zooxanthellae growth, reproduction, etc... Although nitrogen can also be acquired through particulate feeding when not available in its inorganic forms, not all coral species are able to efficiently graze on zooplankton [101] or increase their grazing rates when they are under thermal stress [23]. Such species will therefore be the first to suffer from thermal stress. A short-term seawater acidification alone does not seem to induce strong changes in the capacity of corals to take up nutrients; however, longer-term experiments are still needed to confirm these first results and the lack of effect. Nevertheless, when combined to high temperatures, a decrease in seawater pH had in turn a negative impact on phosphate uptake. Since phosphorus is needed in several steps of the photosynthetic machinery [102,103] as well as in energy production (ATP for example), this study adds to the growing body of evidence that corals will suffer from global change.

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Author Contributions

Conceived and designed the experiments: CG FH RG CFP. Performed the experiments: CG RG CFP. Analyzed the data: CG. Contributed reagents/materials/analysis tools: CG FH RG CFP. Wrote the paper: CG CFP.

References

1. Kleypas JA, Buddemeier RW, Archer D, Gattuso J-P, Langdon C, et al. (1999) Geochemical consequences of increased atmospheric carbon dioxide on coral reefs. *Science* 284: 118–120.
2. Solomon S, Qin D, Manning M, Marquis M, Averyt K, et al. (2007) Climate change 2007: the physical science basis. New York: Cambridge Univ. Press. Intergovernmental Panel on Climate Change. 2007 p.

3. Raven J, Caldeira K, Elderfield H, Hoegh-Guldberg O, Liss P, et al. (2005) Ocean acidification due to increasing atmospheric carbon dioxide. London: The Royal Society.
4. Orr JO, Fabry VJ, Aumont O, Bopp L, Doney SC, et al. (2005) Anthropogenic ocean acidification over the twenty-first century and its impact on calcifying organisms. *Nature* 437: 681–686.
5. Feely RA, Sabine CL, Lee K, Berelson W, Kleypas J, et al. (2004) Impact of Anthropogenic CO₂ on the CaCO₃ System in the Oceans. *Science* 305: 362–366.
6. Caldeira K, Wickett ME (2003) Oceanography: Anthropogenic carbon and ocean pH. *Nature* 425: 365.
7. Hoegh-Guldberg O, Mumby PJ, Hooten AJ, Steneck RS, Greenfield P, et al. (2007) Coral reefs under rapid climate change and ocean acidification. *Science* 318: 1737–1742.
8. Langdon C, Atkinson MJ (2005) Effect of elevated pCO₂ on photosynthesis and calcification of corals and interactions with seasonal change in temperature/irradiance and nutrient enrichment. *J Geophys Res* 110: 1–16.
9. Reynaud S, Leclercq N, Romaine-Lioud S, Ferrier-Pagès C, Jaubert J, et al. (2003) Interacting effects of CO₂ partial pressure and temperature on photosynthesis and calcification in a scleractinian coral. *Glob Change Biol* 9: 1660–1668.
10. Guinotte JM, Fabry VJ (2008) Ocean Acidification and Its Potential Effects on Marine Ecosystems. *Ann N Y Acad Sci* 1134: pp 320–342.
11. Beman JM, Chow C-E, King AL, Feng Y, Fuhrman JA, et al. (2011) Global declines in oceanic nitrification rates as a consequence of ocean acidification. *Proc Natl Acad Sci* 108: 208–213.
12. Hutchins DA, Fu F-X (2008) Linking the oceanographic biogeochemistry of iron and phosphorus with the marine nitrogen cycle. In: Capone DG, Bronk DA, Mulholland MR, Carpenter EJ, eds. Nitrogen in the marine environment, 2nd ed. Amsterdam: Elsevier Press. pp 1,627–621, 653.
13. Levitan O, Rosenberg G, Setlik I, Setlikova E, Grigel J, et al. (2007) Elevated CO₂ enhances nitrogen fixation and growth in the marine cyanobacterium *Trichodesmium*. *Glob Change Biol* 13: 531–538.
14. Hutchins DA, Fu F-X, Zhang Y, Warner ME, Feng Y, et al. (2007) CO₂ control of *Trichodesmium* N₂ fixation, photosynthesis, growth rates, and elemental ratios: implications for past, present, and future ocean biogeochemistry. *Limnol Oceanogr* 52: 1293–1304.
15. Crossland CJ, Barnes DJ (1983) Dissolved nutrients and organic particulates in water flowing over coral reefs at Lizard Island. *Aust J Mar Freshw Res* 34: 835–844.
16. Muscatine L (1990) The role of symbiotic algae in carbon and energy flux in reef corals. In: Dubinski Z, ed. Coral Reefs. Amsterdam/Pays-Bas: Elsevier. pp 75–87.
17. Muscatine L, D'Elia CF (1978) The uptake, retention and release of ammonium by reef corals. *Limnol Oceanogr* 23: 725–734.
18. Roberts JM, Davies PS, Fixter LM, Preston T (1999) Primary site and initial products of ammonium assimilation in the symbiotic sea anemone *Anemonia viridis*. *Mar Biol* 135: 223–236.
19. Bythell JC (1988) A total nitrogen and carbon budget for the elkhorn coral *Acropora palmata* (Lamarck). *Proc 6th Int Coral Reef Symp*. pp 535–540.
20. Hoegh-Guldberg O, Williamson J (1999) Availability of two forms of dissolved nitrogen to the coral *Pocillopora damicornis* and its symbiotic zooxanthellae. *Mar Biol* 133: 561–570.
21. D'Elia CF (1977) The uptake and release of dissolved phosphorus by reef corals. *Limnol Oceanogr* 22: 301–315.
22. Ferrier-Pagès C, Gattuso JP, Dallot S, Jaubert J (2000) Effect of nutrient enrichment on growth and photosynthesis of the zooxanthellate coral *Stylophora pistillata*. *Coral Reefs* 19: 103–113.
23. Ferrier-Pagès C, Rottier C, Béraud E, Levy O (2010) Experimental assessment of the feeding effort of three scleractinian coral species during a thermal stress: Effect on the rates of photosynthesis. *J Exp Mar Biol Ecol* 390: 118–124.
24. Borell EM, Yuliantri AR, Bischof K, Richter C (2008) The effect of heterotrophy on photosynthesis and tissue composition of two scleractinian corals under elevated temperature. *J Exp Mar Biol Ecol* 364: 116–123.
25. Rahav O, Dubinsky Z, Achiuv Y, Falkowski PG (1989) Ammonium metabolism in the zooxanthellate coral, *Stylophora pistillata*. *P Roy Soc Lond B Bio* 236: 325–337.
26. D'Elia CF, Domotor SL, Webb KL (1983) Nutrient uptake kinetics of freshly isolated zooxanthellae. *Mar Biol* 75: 157–167.
27. Fitt WK, Rees TAV, Yellowlees D (1995) Relationship between pH and the availability of dissolved inorganic nitrogen in the zooxanthella-giant clam symbiosis. *Limnol Oceanogr* 40: 976–982.
28. Fu F-X, Warner ME, Zhang Y, Feng Y, Hutchins DA (2007) Effects of increased temperature and CO₂ on photosynthesis, growth, and elemental ratios in marine *Synechococcus* and *Prochlorococcus* (Cyanobacteria). *J Phycol* 43: 485–496.
29. Conroy JP, Milham PJ, Reed ML, Barlow EW (1990) Increases in phosphorus requirements for CO₂-enriched pine species. *Plant Physiology* 92: 977–982.
30. Cotrufo MF, Ineson P, Scott A (1998) Elevated CO₂ reduces the nitrogen concentration of plant tissues. *Glob Change Biol* 4: 43–54.
31. Rogers GS, Payne L, Milham P, Conroy J (1993) Nitrogen and phosphorus requirements of cotton and wheat under changing atmospheric CO₂ concentrations. *Plant Soil* 155–156: 231–234.
32. Cohen AL, Holcomb M (2009) Why corals care about ocean acidification. *Uncovering the mechanism*. *Oceanography* 22: 118–127.
33. Erez J, Reynaud S, Silverman J, Schneider K, Allemand D (2011) Coral calcification under ocean acidification and global change. In: Dubinsky Z, Stambler N, eds. Coral Reefs: An Ecosystem in Transition: Springer Netherlands. pp 151–176.
34. Godinot C, Grover R, Allemand D, Ferrier-Pagès C (2011) High phosphate uptake requirements of the scleractinian coral *Stylophora pistillata*. *J Exp Biol* 214: 2749–2754.
35. Kleypas JA, Danabasoglu G, Lough JM (2008) Potential role of the ocean thermostat in determining regional differences in coral reef bleaching events. *Geophys Res Lett* 35: L03613.
36. Hoegh-Guldberg O (1999) Climate change, coral bleaching and the future of the world's coral reefs. *Mar Freshw Res* 50: 839–866.
37. McClanahan TR, Ateweberhan M, Graham NAJ, Wilson SK, Ruiz Sebastian C, et al. (2007) Western Indian Ocean coral communities: bleaching responses and susceptibility to extinction. *Mar Ecol Prog Ser* 337: 1–13.
38. Kleypas JA, Buddemeier RW, Gattuso JP (2001) The future of coral reefs in an age of global change. *Int J Earth Sci* 90: 426–437.
39. Guinotte J, Buddemeier R, Kleypas J (2003) Future coral reef habitat marginality: temporal and spatial effects of climate change in the Pacific basin. *Coral Reefs* 22: 551–558.
40. Coles SL, Brown BE (2003) Coral Bleaching - Capacity for acclimatization and adaptation. *Adv Mar Biol* 46: 182–223.
41. Hoegh-Guldberg O, Smith GJ (1989) The effect of sudden changes in temperature, light and salinity on the population density and export of zooxanthellae from the reef corals *Stylophora pistillata* Esper and *Seriatopora lysteri* Dana. *J Exp Mar Biol Ecol* 129: 279–303.
42. Jokiel PL, Coles SL (1990) Response of hawaiian and other indo-pacific reef corals to elevated temperature. *Coral Reefs* 8: 155–162.
43. Lesser MP (1997) Oxidative stress causes coral bleaching during exposure to elevated temperatures. *Coral Reefs* 16: 187–192.
44. Cantin NE, Cohen AL, Karnauskas KB, Tarrant AM, McCorkle DC (2010) Ocean warming slows coral growth in the central Red Sea. *Science* 329: 322–325.
45. Lough JM, Barnes DJ (2000) Environmental controls on growth of the massive coral *Porites*. *J Exp Mar Biol Ecol* 245: 225–243.
46. Iglesias-Prieto R, Matta JL, Robins WA, Trench RK (1992) Photosynthetic response to elevated temperature in the symbiotic dinoflagellate *Symbiodinium microadriaticum* in culture. *Proc Natl Acad Sci USA* 89: 10302–10305.
47. Al-Horani F (2005) Effects of changing seawater temperature on photosynthesis and calcification in the scleractinian coral *Galaxea fascicularis*, measured with O₂, Ca²⁺ and pH microsensors. *Sci Mar* 69: 347–354.
48. Anthony KRN, Kline DI, Diaz-Pulido G, Dove S, Hoegh-Guldberg O (2008) Ocean acidification causes bleaching and productivity loss in coral reef builders. *Proc Natl Acad Sci USA* 105: 17442–17446.
49. Tambuté É, Allemand D, Jaubert J (1995) The *Stylophora pistillata* microcolony: a model for studying calcium transport process during coral biomineralization. *Bull Inst océanogr, Monaco, Proceedings of the 7th international symposium on biomineralization, "Biomineralization 93"*, D. Allemand, J PCuif, eds. N° spécial 14: 79–87.
50. Williams GJ, Knapp IS, Maragos JE, Davy SK (2010) Modeling patterns of coral bleaching at a remote Central Pacific atoll. *Mar Pollut Bull* 60: 1467–1476.
51. Lesser MP, Farrell J (2004) Exposure to solar radiation increases damage to both host tissues and algal symbionts of corals during thermal stress. *Coral Reefs* 23: 367–377.
52. Dove S (2004) Scleractinian corals with photoprotective host pigments are hypersensitive to thermal bleaching. *Mar Ecol Prog Ser* 272: 99.
53. Houlbreque F, Rodolfo-Metalpa R, Jeffrey RA, Oberhansli F, Teysse J-L, et al. (2011) Effects of increased pCO₂ on zinc bioaccumulation and calcification in the tropical coral *Stylophora pistillata*. *Coral Reefs* In press.
54. Dickson AG, Sabine CL, Christian JR (2007) Guide to best practices for ocean CO₂ measurements. 8 PspN, editor Sidney, BC: North Pacific Marine Science Organization. 191 p.
55. Godinot C, Ferrier-Pagès C, Grover R (2009) Control of phosphate uptake by zooxanthellae and host cells in the scleractinian coral *Stylophora pistillata*. *Limnol Oceanogr* 54: 1627–1633.
56. Holmes RM, Aminot A, Kérouel R, Hooker BA, Peterson BJ (1999) A simple and precise method for measuring ammonium in marine and freshwater ecosystems. *Can J Fish Aquat Sci* 56: 1801–1808.
57. Murphy J, Riley JP (1962) A modified single solution method for the determination of phosphate in natural waters. *Anal Chim Acta* 27: 31–36.
58. Aminot A, Kérouel R (2007) Dosage automatique des nutriments dans les eaux marines: méthodes en flux continu. Ed. Ifremer, Méthodes d'analyse en milieu marin. 188 p.
59. Jeffrey SW, Humphrey GF (1975) New spectrophotometric equations for determining chlorophylls a, b, c1, and c2 in higher plants, algae, and natural phytoplankton. *Biochem Physiol Pflanz* 167: 191–194.
60. Johannes RE, Weibe WJ (1970) A method for determination of coral tissue biomass and composition. *Limnol Oceanogr* 15: 822–824.
61. Fitt WK, Warner ME (1995) Bleaching patterns of four species of Caribbean reef corals. *Biol Bull* 189: 298–307.

62. Kajiwara K, Nagai A, Ueno S, Yokochi H (1995) Examination of the effect of temperature, light intensity and zooxanthellae concentration on calcification and photosynthesis of scleractinian coral *Acropora pulchra*. *J School Mar Sci Technol* 40: 95–103.
63. Jones RJ, Hoegh-Guldberg O, Larkum AWD, Schreiber U (1998) Temperature-induced bleaching of corals begins with impairment of the CO₂ fixation mechanism in zooxanthellae. *Plant Cell Environ* 21: 1219–1230.
64. Coles SL, Jokiel PL (1977) Effects of temperature on photosynthesis and respiration in hermatypic corals. *Mar Biol* 43: 209–216.
65. Warner ME, Fitt WK, Schmidt GW (1996) The effects of elevated temperature on the photosynthetic efficiency of zooxanthellae in hospite from four different species of reef coral: a novel approach. *Plant Cell Environ* 19: 291–299.
66. Iglesias-Prieto R (1995) The effects of elevated temperature on the photosynthetic response of symbiotic dinoflagellates. *Photosynthesis from light to biosphere* Mathis P, ed. Netherland: Kluwer Academic Publishers Vol 4: 793–796.
67. Schneider K, Erez J (2006) The effect of carbonate chemistry on calcification and photosynthesis in the hermatypic coral *Acropora eurystroma*. *Limnol Oceanogr* 51: 1284–1293.
68. Leclercq N, Gattuso J-P, Jaubert J (2002) Primary production, respiration, and calcification of a coral reef mesocosm under increased CO₂ partial pressure. *Limnol Oceanogr* 47: 558–564.
69. Langdon C, Broecker W, Hammond D, Glenn E, Fitzsimmons K, et al. (2003) Effect of elevated CO₂ on the community metabolism of an experimental coral reef. *Global Biogeochem Cy* 17: 11.11–11.14.
70. Burris JE, Porter JW, Laing WA (1983) Effects of carbon dioxide concentration on coral photosynthesis. *Mar Biol* 75: 113–116.
71. Herfort L, Thake B, Taubner I (2008) Bicarbonate stimulation of calcification and photosynthesis in two hermatypic corals. *J Phycol* 44: 91–98.
72. Al-Moghrabi S, Goiran C, Allemand D, Speziale N, Jaubert J (1996) Inorganic carbon uptake for photosynthesis by the symbiotic coral/dinoflagellate association. II. Mechanisms for bicarbonate uptake. *J Exp Mar Biol Ecol* 199: 227–248.
73. Goiran C, Al-Moghrabi S, Allemand D, Jaubert J (1996) Inorganic carbon uptake for photosynthesis by the symbiotic coral/dinoflagellate association. I. Photosynthetic performances of symbionts and dependence on sea water bicarbonate. *J Exp Mar Biol Ecol* 199: 207–225.
74. Hii Y-S, Bolong AMA, Yang T-T, Liew H-C (2009) Effect of elevated carbon dioxide on two scleractinian corals: *Porites cylindrica* (Dana, 1846) and *Galaxea fascicularis* (Linnaeus, 1767). *J Mar Biol* 2009: pp 1–7.
75. Lehninger AL (1985) Enzymes. In: Lehninger AL, Flammarion Medecine Sciences, eds. Principles of biochemistry, 1st Edition. Baltimore: The John Hopkins University. pp 207–248.
76. Fersht A (1977) Enzyme structure and mechanism. San Francisco: W. H. Freeman.
77. D'Elia CF, Cook CB (1988) Methylamine uptake by zooxanthellae-invertebrate symbioses: insights into host ammonium environment and nutrition. *Limnol Oceanogr* 33: 1153–1165.
78. Wilkerson FP, Trench RK (1986) Uptake of dissolved inorganic nitrogen by the symbiotic clam *Tridacna gigas* and the coral *Acropora* sp. *Mar Biol* 93: 237–246.
79. Domotor SL, D'Elia CF (1984) Nutrient uptake kinetics and growth of zooxanthellae maintained in laboratory culture. *Mar Biol* 80: 93–101.
80. Wilkerson F, Muscatine L (1984) Uptake and assimilation of dissolved inorganic nitrogen by a symbiotic sea anemone. *Proc R Soc Lond B Biol Sci B*, pp 71–86.
81. Cui L, Cao R, Li J, Zhang L, Wang J (2006) High temperature effects on ammonium assimilation in leaves of two *Festuca arundinacea* cultivars with different heat susceptibility. *Plant Growth Regul* 49: 127–136.
82. Yellowlees D, Rees TAV, Fitt WK (1994) Effect of ammonium-supplemented seawater on glutamine synthetase and glutamate deshydrogenase activities in host tissue and zooxanthellae of *Pocillopora damicornis* and on ammonium uptake rates of the zooxanthellae. *Pac Sci* 48: 291–295.
83. Anderson LA, Burris JE (1987) Role of glutamine synthetase in ammonia assimilation by symbiotic marine dinoflagellates (zooxanthellae). *Mar Biol* 94: 451–458.
84. Catmull J, Yellowlees D, Miller DJ (1987) NADP⁺-dependent glutamate dehydrogenase from *Acropora formosa*: purification and properties. *Mar Biol* 95: 559–563.
85. Marubini F, Atkinson MJ (1999) Effects of lowered pH and elevated nitrate on coral calcification. *Mar Ecol Prog Ser* 188: 117–121.
86. Ertan H (1992) Some properties of the NADPH-dependent glutamate synthetase and glutamate synthase from *Corynebacterium callunae*. *Arch Microbiol* 158: 35–41.
87. Martinez-Bilbao M, Martinez A, Urkijo I, Llama MJ, Serra JL (1988) Induction, isolation, and some properties of the NADPH-dependent glutamate dehydrogenase from the nonheterocystous cyanobacterium *Phormidium laminosum*. *J Bacteriol* 170: 4897–4902.
88. Grover R, Maguer J-F, Reynaud-Vaganay S, Ferrier-Pagès C (2002) Uptake of ammonium by the scleractinian coral *Stylophora pistillata*: Effect of feeding, light, and ammonium concentrations. *Limnol Oceanogr* 47: 782–790.
89. Toetz DW (1981) Effects of pH, phosphate and ammonia on the rate of uptake of nitrate and ammonia by freshwater phytoplankton. *Hydrobiologia* 76: 23–26.
90. Xu Z, Zou D, Gao K (2010) Effects of elevated CO₂ and phosphorus supply on growth, photosynthesis and nutrient uptake in the marine macroalga *Gracilaria lemaneiformis* (Rhodophyta). *Bot Mar* 53: 123–129.
91. Zou D (2005) Effects of elevated atmospheric CO₂ on growth, photosynthesis and nitrogen metabolism in the economic brown seaweed, *Hizikia fusiforme* (Sargassaceae, Phaeophyta). *Aquaculture* 250: 726–735.
92. Gordillo E JL, Niell FX, Figueroa FL (2001) Non-photosynthetic enhancement of growth by high CO₂ level in the nitrophilic seaweed *Ulva rigida* C. Agardh (Chlorophyta). *Planta* 213: 64–70.
93. Grover R, Maguer J-F, Allemand D, Ferrier-Pagès C (2003) Nitrate uptake in the scleractinian coral *Stylophora pistillata*. *Limnol Oceanogr* 48: 2266–2274.
94. Grover R, Maguer JF, Allemand D, Ferrier-Pagès C (2006) Urea uptake by the scleractinian coral *Stylophora pistillata*. *J Exp Mar Biol Ecol* 332: 216–225.
95. Jackson AE, Yellowlees D (1990) Phosphate uptake by zooxanthellae isolated from corals. *Proc R Soc Lond B Biol Sci* 242: 201–204.
96. Sorokin YI (1992) Phosphorus metabolism in coral reef communities: exchange between the water column and bottom biotopes. *Hydrobiologia* 242: 105–114.
97. Tanaka T, Thingstad TF, Lovdal T, Grossart HP, Larsen A, et al. (2008) Availability of phosphate for phytoplankton and bacteria and of glucose for bacteria at different pCO₂ levels in a mesocosm study. *Biogeosciences* 5: 669–678.
98. Gerencser GA, Levin R, Zhang J (2002) Sodium-phosphate symport by *Aphysia Californica* gut (physiology). *Zool Sci (Tokyo)* 19: 163–166.
99. Markovich D (2010) Sulfate and phosphate transporters in mammalian renal and gastrointestinal systems. In: Gerencser GA, ed. Epithelial transport physiology. New York: Springer.
100. Virkki LV, Biber J, Murer H, Forster IC (2007) Phosphate transporters: a tale of two solute carriers. *Am J Physiol Renal Physiol* 293: 643–654.
101. Grottoli AG, Rodrigues LJ, Palardy JE (2006) Heterotrophic plasticity and resilience in bleached corals. *Nature*. pp 1186–1189.
102. Goodman M, Bradley DF, Calvin M (1953) Phosphorus and photosynthesis. I. Differences in the light and dark incorporation of radiophosphate. *J Am Chem Soc* 75: 1962–1967.
103. Hall DO, Rao KK (1987) Photosynthesis. Cambridge: Cambridge University Press. 128 p.